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Socially informed dispersal in a territorial cooperative breeder

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ABSTRACT

1. Dispersal is a key process governing the dynamics of socially and spatially structured populations, and involves three distinct stages: emigration, transience, and settlement. At each stage, individuals have to make movement decisions, which are influenced by social, environmental, and individual factors. Yet, a comprehensive understanding of the drivers that influence such decisions is still lacking, particularly for the transient stage during which free-living individuals are inherently difficult to follow.
2. Social circumstances such as the likelihood of encountering conspecifics can be expected to strongly affect decision making during dispersal, particularly in territorial species where encounters with resident conspecifics are antagonistic. Here we analyzed the movement trajectories of 47 dispersing coalitions of Kalahari meerkats (*Suricata suricatta*) through a landscape occupied by constantly monitored resident groups, while simultaneously taking into account environmental and individual characteristics.
3. We used GPS locations collected on resident groups to create a geo-referenced social landscape representing the likelihood of encountering resident groups. We used a step-selection function to infer the effect of social, environmental and individual covariates on habitat selection during dispersal. Lastly, we created a temporal mismatch between the social landscape and the dispersal event of interest to identify the temporal scale at which dispersers perceive the social landscape.
4. Including information about the social landscape considerably improved our representation of the dispersal trajectory, compared to analyses that only accounted for environmental variables. The latter were only marginally selected or avoided by dispersers. Before leaving their natal territory, dispersers selected areas frequently used by their natal group. In contrast, after leaving their natal territory, they selectively used areas where they

were less likely to encounter unrelated groups. This pattern was particularly marked in larger dispersing coalitions and when unrelated males were part of the dispersing coalition.

5. Our results suggest that, in socially and spatially structured species, dispersers gather and process social information during dispersal, and that reducing risk of aggression from unrelated resident groups outweighs benefits derived from conspecific attraction. Finally, our work underlines the intimate link between the social structure of a population and dispersal, which affect each other reciprocally.

Keywords: Conspecific avoidance, informed dispersal, movement, social landscape, step selection, *Suricata suricatta*

INTRODUCTION

Dispersal of individuals is an important process governing the dynamics of spatially and socially structured populations (Hanski 1999; Clobert *et al.* 2001). Through emigration and immigration, dispersal can also affect local allele frequencies, alter the structure of existing social groups, and lead to the formation of new groups (Clobert *et al.* 2001; Bateman *et al.* 2012). Improving our understanding of dispersal is therefore critical as we strive for better forecasting of long-term population trends and viability (Bowler & Benton 2005; Ronce 2007).

The dispersal process can be divided into three phases: emigration from the natal territory; transience; and settlement in a new territory (Bowler & Benton 2005). Each phase is dependent on different individual, environmental, and social drivers, and our ability to understand the mechanisms of dispersal relies on our ability to account for and understand the factors affecting the movement of individuals in each of the three phases (Wiens 2001; Schick *et al.* 2008; Clobert *et al.* 2009). Due to the difficulty of following individuals under natural conditions, past studies have mainly focused on the causes and consequences of emigration from the natal territory (Stephens *et al.* 2005; Cote & Clobert 2007, 2010). For instance, rates of emigration have been shown to be density-dependent, to relax local competition, and to favour inbreeding avoidance (Holekamp & Sherman 1989; Perrin & Mazalov 2000; Bowler & Benton 2005; Bateman *et al.* 2012). In contrast, we have much less information about the mechanisms and strategies that individuals use during the transient and settlement phases of dispersal (Travis *et al.* 2012; Elliot *et al.* 2014). Empirical studies that focused on the transient phase have concentrated on changes in movement characteristics and on the relationship between movement trajectories and the surrounding environmental landscape. The combined influence of the distribution and abundance of conspecifics – hereafter referred to as social landscape – on decision making during transience has received relatively little empirical attention, and related inferences are often based solely on theoretical work (Fletcher 2006;

Travis *et al.* 2012; Gilroy & Lockwood 2016). Yet the social landscape is likely to exert an important influence on the behaviour of dispersing individuals during the transient phase.

In territorial social species such as the lion (*Panthera leo*), the African wild dog (*Lycaon pictus*) and the meerkat (*Suricata suricatta*), encounters with unrelated groups and individuals are often antagonistic and costly particularly for the smaller party (McComb, Packer & Pusey 1994; Creel & Creel 2002; Mares *et al.* 2011). Dispersers are typically outnumbered by resident groups, and thus, they may avoid areas where encounters with resident groups are likely to occur. Such avoidance may be particularly strong in smaller dispersing coalitions, and shortly before settlement, when dispersers seek an exclusive territory. At the opposite end, the search for potential mates and high quality habitats during transience and settlement may bring dispersers closer to resident groups (Andreassen & Ims 2001; Glorvigen *et al.* 2012). Whether dispersers are attracted to resident groups is likely to be related to whether or not they can ever join them. Thus, dispersing male meerkats would be expected to be attracted to resident groups, females to avoid them (Stephens *et al.* 2005; Mares *et al.* 2011). In species that aggressively defend their territory from unrelated conspecifics and where encounters can have fatal consequences, reducing costs of aggression may outweigh the benefits derived from conspecific attraction (Stamps 2001; Fletcher 2006). Irrespective of which strategy (conspecific avoidance or attraction) dispersers adopt during transience, their ability to assess the social landscape at different spatial and temporal scales can reduce costs and increase dispersal success (Bonte *et al.* 2012). Scent marks and other signs left by residents convey information about their distribution and status (Jordan, Cherry & Manser 2007; Mares *et al.* 2011; Jackson, McNutt & Apps 2012) and may be used by dispersing individuals to indirectly gain information of the social landscape through which they move, avoiding risky direct encounters.

A long-term study of the Kalahari meerkat (*Suricata suricatta*) (Clutton-Brock & Manser 2016) provides a unique opportunity to investigate the role of the social and environmental landscape during the transient phase of dispersal in a social and territorial species. Meerkats are territorial cooperative breeders living in groups of 5-40 individuals, and each group occupies an exclusive territory of 1-5 km² (Clutton-Brock *et al.* 1999; Jordan *et al.* 2007). Intruding meerkats, being dispersers or individuals from neighbouring groups, are readily challenged and chased by territorial groups, and such clashes can have severe consequences (Mares, Young & Clutton-Brock 2012). Groups are characterized by the presence of a dominant pair that monopolizes most of reproduction; subordinate individuals delay dispersal and help rearing their siblings. Older subordinate females are aggressively evicted from their natal group, and they form same-sex dispersing coalitions (Clutton-Brock *et al.* 1998). Encounters with the natal group after eviction typically result in aggressive behaviour directed towards the evicted individuals (Young *et al.* 2006). Dispersing coalitions either leave the natal territory and search for unrelated males and a place to settle, or they are sometimes let to re-join their natal group after the dominant female has given birth (Clutton-Brock *et al.* 1998). Subordinate females are, however, typically evicted in successive breeding attempts by the dominant female and eventually leave the area to form a new group (Clutton-Brock *et al.* 1998). Evictions happen year-round, but typically peaks between September and March (hereafter referred to as the dispersal season). Dispersing coalitions may have to travel a considerable distance to locate vacant territories and during this period experience costs of increased predation pressure and aggression from other meerkat groups (Clutton-Brock *et al.* 1999). In contrast to females, subordinate male meerkats spontaneously leave their natal group and prospect for extra-group mating opportunities. Such prospecting forays typically last few days, and a small proportion of them culminates in the males' dispersal or disappearance

(Young, Spong & Clutton-Brock 2007), likely after encountering a coalition of unrelated dispersing females.

In this study, we aimed to investigate the effect of the social landscape on movement behaviour and decision making during the transient phase of dispersal in a socially and spatially structured species. To do so, we fitted GPS radio collars on dispersing coalitions of Kalahari meerkats and followed their movements across territories occupied by constantly monitored resident groups. We created environmental maps from satellite imagery to control for the effect of habitat types, taken to represent different habitat quality, on movements. We used step-selection models to quantify habitat and social selection of dispersers before and after they left their natal territory. We regularly visited dispersers and collected information on coalition composition to evaluate the effect of coalition size on selection of locations in the social landscape during transience. Finally, we identified the temporal scale at which dispersing coalitions perceived and reacted to the social landscape and investigated whether they integrated information about the distribution of resident groups collected over different weeks and months.

METHODS

Study system and data collection

The study site was located at the Kuruman River Reserve (S 26.92289°, E 21.84048) and the adjacent ranch-land near Van Zylsrus, South Africa. The site includes a section of the fossil Kuruman riverbed; elsewhere, the landscape is typical of the Kalahari Desert and is characterized by sparsely vegetated sand dunes and dry pans (Fig. 1A). The climate is characterized by a dry season between April and November and a wet season between December and March, with 250 mm annual average precipitation. For further details on the study site see Mares *et al.* (2012) and Young (2003).

The local population of wild meerkats has been subject of a long-term study started in 1993, and all study animals were individually marked. In any given year, between 15 and 18 resident groups that inhabit an area of 50-60 km² were regularly monitored. Resident meerkat groups were habituated to the close presence (< 1 m) of researchers who visited the groups every two to four days to collect behavioural and life-history data. During group visits, which took place early mornings and late afternoons, and lasted three to four hours, researchers consistently collected GPS locations every 15 minutes using a handheld GPS unit. We used this spatial information on resident groups to create social landscapes (further detailed below).

We mounted GPS radio collars on subordinate dispersing females few days prior to or immediately after eviction. We used behavioural cues such as aggression from the dominant female, restless behaviour, and social withdrawal to identify potential disperser and anticipate the time of eviction. We captured and anesthetized meerkats following a standardized protocol used at the Kalahari Meerkat Project (Jordan *et al.* 2007). All captures were approved by the South African Department of Environment and Nature Conservation and were performed under permit 'FAUNA 192/2014'.

The GPS radio collars were composed of a 3.2 g stand-alone VHF beacon module (Holohil Systems Ltd., Canada) and a 16.0 g stand-alone GPS module with integrated ZigBee bidirectional communication capability powered by a 3.6 V ½ AA lithium battery (CDD Ltd., Greece). We casted the two modules in epoxy resins and used a 4.5 mm wide cable tie coated with soft heat-shrink tubing to attach the unit to the neck of the animals. Overall collar weight was < 25 g, corresponding to 3-4 % of the animal body weight. Collars of this size and weight do not affect meerkats (Golabek, Jordan & Clutton-Brock 2008). Only one individual was equipped with a GPS radio collar in each dispersing coalition at any given time. We programmed the collars to collect hourly GPS locations between sunrise and sunset as meerkats sleep underground at night. We additionally fitted few individuals with VHF-only

radio collars (Biotrack, United Kingdom) weighing approximately 23 g, when deploying GPS collars was not possible for logistic reasons. For coalitions fitted with VHF-only collars, we manually collected GPS information at each visit (Supplementary Material S1).

Dispersing coalitions

We defined a dispersing coalition as a single female or multiple females, evicted from the natal group. We tracked each dispersing coalition by means of VHF telemetry every two to seven days to record coalition size, which included females and, when present, unrelated males. We defined a dispersal event as the collection of events starting at eviction and lasting until either return to the natal group, settlement in a new territory, or death before settlement. We did not include data collected after two weeks past successful settlement in a new territory as coalitions may show resident-like behaviour after settlement.

Within each dispersal season (September – March), we merged consecutive dispersal events by the same dispersing coalition in a single combined dispersal event (Supplementary Material S1). For example, if two sisters were evicted from the resident group, then accepted back into the group, and evicted again shortly after, these two events were considered as one single dispersal event. This procedure reduced data dependence (two dispersal events of the same coalition separated by a short period of time cannot be considered as independent) and increased the data points of each combined dispersal event, which is a prerequisite for fitting robust individual-level models for the step selection function (Craiu, Duchesne & Fortin 2008; Fieberg *et al.* 2010; see below). During a three-year period, we monitored a total of 47 dispersing events, of which five were combined dispersal events consisting of two or three consecutive events (Supplementary Material S1). Unless otherwise specified, all analyses were performed using data from 47 dispersing coalitions.

We collected an average of 137 (range: 10 – 1114) GPS locations per dispersal event, which lasted on average 53 days (Supplementary Material S1). Twenty-seven coalitions re-joined their natal group after a variable duration of one (i.e. a coalition was evicted one day and allowed back to the group the following day) and 76 days. The remaining 20 coalitions did not return to their natal group and either dispersed for a minimum of six and a maximum of 284 days before settling in a new territory (n = 12 coalitions), dying (n = 3) or being lost (n = 5) before settlement.

Creation of geo-referenced environmental layers

Within R 3.2.0 (The R Foundation for Statistical Computing), we used the *rgdal* package to derive a geo-referenced digital elevation map (DEM) from Advanced Spaceborne Thermal Emission and Reflection Radiometer (ASTER) images (Fig. 1B). We further used NASA's Landsat 7 satellite images at 30x30 meters resolution to create a geo-referenced habitat map following an unsupervised classification approach using the function *kmeans* (Fig. 1A). Given the expected difference in reflectance between the vegetated sandy dunes and the dry 'pans', this unsupervised approach was adequate. Following ground-truthed data at 80 randomly selected locations, we merged the five classes (cfr. Fig. 1A) that resulted from the unsupervised classification into three main habitat classes; namely "pans", "sandy matrix", and "transition" between the former two classes. This procedure allowed us to achieve an accuracy of 98% for pans (1 in 45 pan locations was misclassified as transition), of 73% for sandy matrix (four in 15 matrix locations were misclassified as transition), and of 47% for transition. Pans were never misclassified as matrix and vice versa (for further details see Supplementary Material S2). The map obtained following this approach was very similar to vegetation maps for the study site used by (Bateman *et al.* 2015) and Google Earth imagery (Google Inc. 2012) thus confirming the reliability of the method (Fig. 1). Both ASTER and Landsat 7 images used are

available through the United States Geological Survey agency (<http://www.usgs.gov>).

Additionally, we used Google Earth to manually digitize and export the main channel of the Kuruman River (Fig. 1).

Creation of geo-referenced social layers

We used GPS location data collected on resident groups to create a geo-referenced social landscape (Fig. 1C). Because GPS locations on resident groups were collected every 15 minutes throughout a visit and were therefore highly spatially and temporally correlated, we computed the harmonic mean of all locations collected during one visit resulting in one GPS location per visit. For each resident group, we created a kernel utilization density (KUD) map with the function *kernelUD* in the *adehabitatHR* package (Calenge 2006). We reduced the reference smoothing parameter h_{ref} until the smallest possible contiguous territory (i.e., one single polygon) was obtained (Kie *et al.* 2010; Fieberg & Börger 2012). This procedure was required because h_{ref} typically over-smooths the data, inflating territory size. We used GPS data collected during a time interval of 30 days prior to eviction of the dispersing coalition of interest to create the KUD maps for each resident group. We excluded resident groups with less than 10 locations collected over the 30-day interval to reduce the likelihood of unreliable KUD estimation. We multiplied KUD pixel values by group size and summed KUD raster maps of each resident group to create a contiguous social landscape comprising the distribution and abundance of all resident groups for the given 30-day interval. Low pixel values represented non-risky area characterized by a small likelihood of encountering resident groups, while high pixel values represented risky areas. We assigned NA values to pixels with no information on resident groups. For those dispersing events lasting more than two weeks, we recalculated the social landscape with a two-week sliding window. This procedure allowed us to create a dynamic social landscape. A social landscape was created using data on resident

groups collected within 30 days prior to eviction. The dispersing coalition of interest was then ‘allowed to’ move through this social landscape for two weeks (days 0 to 15 of the dispersal trajectory). If the dispersal event lasted more than two weeks, a second social landscape was created using data collected on resident individuals within 30 days prior to day 15 of the dispersal event. The same dispersing coalition was then ‘allowed to’ move through this new social landscape for two weeks (days 16 to 30 of the dispersal trajectory); and so forth (see Supplementary Material S3 for a dynamic representation of the social landscape).

To investigate at what temporal scales dispersing individuals perceive the social landscape, we repeated the same process for alternative time intervals of 45, 60, 75, 90, 105, 120, 150, 180, 240 days. In a final analysis, which we used as control treatment, we created a temporal mismatch between the social landscape and the dispersal events. For this, we created the social landscape by shifting back the beginning of each time interval by 180 days compared to the time of eviction of the dispersing coalition of interest. Because we assumed such a social landscape to be of little relevance for dispersers, we expected no consistency in their selection and avoidance of presumed risky areas.

Statistical Analysis

We used a step-selection function (SSF) to infer the effect of social and environmental covariates on habitat selection during dispersal (Fortin *et al.* 2005). Within this framework, the decision of a coalition to use a given location is conditioned on environmental and social characteristics at alternative locations (Fortin *et al.* 2009) that the coalition could have reached within an hour (the sampling interval between consecutive relocations). For each observed (i.e. used) step along a dispersing coalition path, we created ten alternative steps; the end of these steps represented alternative locations that the coalition could have chosen. A step is here defined as the vector between two consecutive relocations spaced one hour apart. We created

alternative steps by randomly drawing angles from a continuous uniform distribution $U(0, 2\pi)$ and steps from a normal distribution $N(\mu, \sigma)$ with μ equal to two times the mean empirical step length of monitored dispersing coalitions, with σ equal to one step length standard deviation. We chose μ to allow alternative relocations to be far from the relocation used by the focal coalition, and thus to be characterized by different environmental and social factors. Dispersing meerkats are bound to sleeping burrows, with new burrows being used as the dispersal process progresses. While steps between consecutive burrows can be expected to be directional (i.e. small turn angles), steps in the vicinity of sleeping holes are more circuitous (i.e. equally distributed turn angles). As we could not distinguish between these two types, drawing angles from a uniform distribution appeared more parsimonious.

Following this matched case-control design (Fortin *et al.* 2005), we built SSF using conditional logistic regression to compare used (scored 1) and alternative (scored 0) locations. The SSF took the general form

$$w(x) = \exp(\beta_1 x_1 + \beta_2 x_2 + \dots + \beta_{n-2} x_{n-2} + \beta_{n-1} x_1 : x_2 + \beta_n x_1 : x_3)$$

where β 's are selection coefficients associated with covariates and biologically meaningful interaction terms (Fortin & Fortin 2009). Steps with higher SSF scores $w(X)$ are more likely to be chosen by the animals (Fortin *et al.* 2005), and $\beta = 0$ indicates absence of selection/avoidance (Forester, Im & Rathouz 2009). The environmental covariates considered in the analyses included ELEVATION (continuous), DISTANCE TO RIVER (continuous), and the binary habitat class PAN (coded 1 if a dispersing coalition was on a pan and 0 if outside). We did not include the other two habitat classes (sandy matrix and transition) to avoid collinearity with pans. The social variables were RISKINESS (the continuous social landscape values), and NATAL (coded 1 if a dispersing coalition was within and 0 if outside the territory boundary of the natal group). Because we anticipated that dispersing coalitions behave differently within and outside the territory of the natal group, we included a RISKINESS by

NATAL interaction in our model. We also added dispersing COALITION SIZE as part of an interaction term with RISKINESS to test whether coalitions of different size reacted differently to the social landscape. Because each used and its ten alternative locations were assigned the same coalition size, coalition size cannot appear in the model as a main effect (Fortin & Fortin 2009). We further defined a CLUSTER as the unique identifier for each independent dispersal event, and a STRATUM as the unique identifier for each combination of used and alternative locations. We used CLUSTER and STRATUM as grouping variables in the analyses (Fortin *et al.* 2005). We implemented the above model in the *clogit* function in the *survival* package to obtain population-level estimates of the selection coefficients associated with each covariate. We calculated alternative SSF models with RISKINESS values obtained at each time intervals (30, 45, ..., 240 days) to detect the temporal scale at which dispersing coalitions showed the strongest response to the social landscape (cfr. Fig. 3). We reported β coefficients associated with all covariates only for the time interval that showed the strongest response.

In a second step, we used β coefficients obtained from conditional logistic regression fitted to each coalition independently (Fieberg *et al.* 2010) to make inference on the effect of additional coalition-specific covariates on step selection. We used linear regression to test the effect of coalition size on estimated β coefficients, which were treated as response variables (Thurfjell, Ciuti & Boyce 2014). Because some coalitions either never left the territory of their natal group or almost only roamed outside of it (Supplementary Material S1), it was not possible to fit a single model including the variable NATAL to each single dispersing coalition. This was due to matrix singularities resulting from a coalition experiencing the same events (e.g. all observed and alternative locations in a stratum have the same NATAL value). We therefore ran two separate models; one with data collected when the dispersing coalitions were within the territory of their natal group, and one when they were outside the natal territory. To obtain robust coalition-specific estimations, we only retained dispersing coalitions

with a minimum of 30 used locations (see Craiu *et al.* 2008). Of the 47 monitored dispersing coalitions, 30 coalitions met this criterion while roaming within the territory of their natal group and 15 while outside (Supplementary Material S1). Like above, we run these two models only for the time window at which RISKINESS showed the strongest response.

In an additional analysis, we used a generalized additive mix model in the *mgcv* package (Wood 2004) to find the shape of any nonlinear relationship of RISKINESS over time (days since eviction) without imposing any specific parametric form. We averaged RISKINESS values on a daily basis to reduce autocorrelation issues. We allowed for differences between coalitions that returned to the natal group and coalitions that settled, while controlling for the effect of coalition size and the presence of unrelated males within a coalition. We included coalitions identity as random term.

RESULTS

Movement patterns during dispersal were largely dependent on coalition size. Of the 47 dispersing coalitions monitored, the coalitions that settled ($n = 12$) were typically larger (mean = 2.3 females or 4.7 females and males) than the coalitions that returned ($n = 33$ coalitions; mean = 1.8 females or 1.9 females and males; Wilcoxon $W = 116$, $p = 0.14$ and $W = 39.5$, $p < 0.01$, for females or females and males respectively) (Supplementary Material S4). Average maximum displacement from the site of capture was 2'263 m (range 434 – 10'742 m). Coalition size showed a positive relationship with the log-transformed maximum distance travelled ($F_{1,46} = 6.14$, $p = 0.017$) after correcting for differences in dispersal time.

After eviction from their natal group, dispersing coalitions selected for low-lying slacks between sand dunes and avoided pans (Table 1). Selection coefficients for PANS in four coalitions appeared implausibly large for an exponential model (e.g. -17.68; Table 2) and were likely due to matrix singularities. We therefore ran two alternative models: in one model, we

excluded these four coalitions, and in the other model we retained all coalitions but removed PANS from the model. Results of coefficients of selection (i.e., β values) remained unchanged. Dispersing coalitions neither selected nor avoided locations closer to (respectively, further away from) the fossil Kuruman riverbed. Dispersing coalitions reacted differently to the social landscape depending on whether they were inside or outside the territory of the natal group (RISKINESS:NATAL interaction term: $\beta = -0.055$, robust SE, 0.014, $p < 0.01$; Table 1). When they were still within the territory of their natal group, they selected for locations characterized by a higher likelihood to encounter the natal group. When they roamed outside the territory of their natal group, they avoided locations characterized by a high likelihood of encountering unrelated resident groups. These results thus support the hypothesis that dispersing coalitions move in a way to minimize interactions with unrelated territorial groups.

Avoidance of risky locations outside the territory of the natal group persisted through all time windows used to determine the social landscape, but the strength of risk avoidance peaked at time windows between 60 and 75 days (Fig. 2). When we introduced a 180-day mismatch between the social landscape and the time of eviction (see methods), we observed no differences in the degree of selection of risky locations for areas inside and outside the territory of the natal group (Fig. 2). This suggested that the social landscape is of relevance only within a time intervals of a few months but is not informative over longer periods. Overall, adding information about the social landscape to the baseline model that only accounted for environmental variables improved model performance (log-likelihood ratio test for nested models: $\chi^2 = 261$, $p < 0.01$, Supplementary Material S5).

Analysis at the coalition level did not support the hypothesis that smaller dispersing coalitions avoided areas where encounters were likely to occur more than larger coalitions. Although not statistically significant, there was an indication that selection for high-likelihood-of-encounter areas decreased as dispersing coalition size increased (Fig. 3); both within (slope

= -0.013, SE = 0.009, $p = 0.14$) and outside (slope = -0.12, SE = 0.11, $p = 0.31$) the territory of the natal group. Overall, results obtained at the coalition level (Table 2) largely supported the results obtained at the population level (Table 1) showing that when dispersing coalitions were within the territory of their natal group, they selected for areas characterized by a higher likelihood to encounter the natal group (Table 2, Fig 3).

As time after eviction progressed, coalitions that settled increasingly used areas where they were less likely to encounter conspecifics ($F = 17.2$, $p < 0.01$), whereas we could not detect any relationship with time for coalitions that eventually re-joined their natal group. The presence of unrelated males was associated with the use of areas characterized by a lower likelihood to encounter resident conspecifics ($t = -3.4$, $p < 0.01$), both for coalitions that re-joined their group and for coalitions that settled elsewhere, while coalition size had no statistically significant effect ($t = -1.1$, $p = 0.3$).

DISCUSSION

Our study showed that, in a population of wild Kalahari meerkats, the movements of dispersing individuals were influenced by the social landscape, i.e. the distribution and abundance of resident conspecifics. Importantly, the reaction to this social landscape, depended on whether dispersers were inside or outside the territory of their natal group. Before leaving their groups' territory, dispersers appeared to maximise their chance of being in the same area as their natal group. After leaving their natal territory, they selectively used areas where they were less likely to encounter other unrelated groups, and increasingly so when unrelated males were part of the dispersing coalition. This contrasting use of the social landscape suggests that dispersing meerkats can acquire information on the distribution and abundance of conspecifics and use this information to minimize antagonistic contacts with unrelated territorial groups.

Seeking proximity of the natal group despite the possibility of aggression, is consistent with existing evidence of the advantages of group living and delayed dispersal in wild Kalahari meerkats (Clutton-Brock *et al.* 1999; Clutton-Brock & Manser 2016). The tendency for dispersing individuals to ‘shadow’ their natal group after being evicted was more pronounced for smaller than for larger dispersing coalitions, suggesting that small coalitions may suffer higher costs or gain smaller benefits by leaving their natal area. Associated costs may be direct costs such as increased mortality due to reduced antipredator vigilance and dilution effect in smaller groups (Delm 1990), or fitness costs associated with reduced recruitment rate in small, newly formed groups (N. Maag pers. comm.).

Outside the territory of the natal group, moving through areas characterized by a small likelihood of encountering unrelated resident conspecifics is likely to be beneficial since it reduces antagonistic interactions and increase dispersal success. In the effort to avoid resident groups, however, dispersers may move through suboptimal and resource-poor areas (Palomares *et al.* 2000). While resident meerkat groups have been shown to concentrate along the edges of the dry riverbed and surrounding flats habitats (Bateman *et al.* 2015), the dispersing coalitions monitored during this study avoided such habitats (here referred to as pans). The riverbed and pans offer more shelters (bolting holes) than the surroundings (Manser & Bell 2004); avoidance of these safer habitats suggests that avoidance of resident groups may be a stronger driver of habitat selection during dispersal than predation pressure. Such strategy may be favoured by a relative low predation pressure; we only recorded one case of predation by a bird of prey (on a single dispersers) and two meerkats were hit by cars. As food availability can mask the effect of predation and conspecific aggression on habitat selection we urge further investigation. In the Kalahari rainfall is highly variable both between and within years and it affects food availability with consequences on meerkats growth and body condition (English, Bateman & Clutton-Brock 2012). Changes in food availability may be partly responsible for

some of the variations in habitat selection observed among individual dispersing coalitions. This in light of the fact that the confidence intervals for the β coefficients of the habitat features considered were only lightly different from zero (which would represent no selection) and hence more susceptible to changes in selection or avoidance following alteration of the characteristics of the habitat itself.

Our results did not support the hypothesis that larger dispersing coalitions are competitively stronger than smaller coalitions and therefore less likely to avoid areas characterized by a high likelihood of encountering unrelated resident groups. This may be due to the fact that no matter how large a dispersing coalition may be, it will inevitably be substantially smaller than a resident group, which averages 15.5 individuals (Bateman *et al.* 2013), and therefore consistently avoid it. Our results suggest that rather than influencing movement patterns during transience, coalition size influenced the final outcome of the dispersal event – with smaller coalitions returning to the natal group and larger coalitions settling in new territories. Movements during transience were instead influenced by the presence of unrelated males and increasing time after eviction. Both factors caused dispersing coalitions to avoid areas with a high likelihood to encounter resident groups, suggesting that newly formed groups seek vacant or little used territories to settle. Males thus not only influence dispersal patterns at emigration and settlement (Cote & Clobert 2007; Davidian *et al.* 2016; Gilroy & Lockwood 2016), but also during the transient phase. In fact, evidence suggests that, in meerkats, the association with unrelated males is one of the key drivers promoting the transition between the consecutive phases of dispersal (N. Maag *et al.* in review).

Results from our study also provide important insight into the temporal scale at which dispersing coalitions perceive and react to the social landscape. The underlying assumption of our analysis is that the GPS locations used to model the distribution of resident groups directly

correlate with cues of their presence and that dispersing coalitions perceive and react to such cues. These may be direct encounters or indirect cues such as scent markings. Our analysis showed strongest avoidance of areas characterized by a high likelihood to encounter unrelated resident groups when the social landscape was modelled using GPS data collected on resident groups during the 60/75-day interval preceding the dispersal trajectory. Using GPS locations collected during shorter or longer time periods resulted in decreasing avoidance. We speculate that a social landscape based on information collected during a too short period (< 30 days) does not accurately represent the risks perceived by dispersing coalitions. The same applies to social landscape based on information collected long before the dispersal event took place, as suggested by our mismatch design (180-days timescale). These results suggest that care must be taken in the temporal offset for the information used to create the social landscape.

The relative roles of direct and indirect cues in shaping dispersers' own representation of the social landscape remains to be determined (Creel *et al.* 2005; Broekhuis *et al.* 2013). The ability to predict risks related to distribution and abundance of conspecifics based on indirect, long-lasting olfactory cues seems a parsimonious mechanistic explanation for a species that relies on scent marking for between- and within-group signalling and communication (Jordan *et al.* 2007; Mares *et al.* 2011; Bateman *et al.* 2015). While we have no direct evidence on how long such cues may last for in meerkats under specific environmental conditions, reaction to olfactory cues of several weeks to months has been documented for mammalian scent marks (20-25 days: *Helogale parvula*; 65 days: sniffer dogs; 40-100 days: *Mesocricetus auratus*; 180 days: *Micricebus coquerely*) (Apps, Weldon & Kramer 2015). Our results suggest similar time windows to be plausible for meerkats, but further investigation is required.

Our work, which focused on understanding the effects of the social landscape on movement behaviour and decision making during the transient phase of dispersal, provides empirical evidence for one of the “four areas where research effort should be directed”, as

identified by Clobert et al. (2009) in their seminal review paper. Our results support the concept of socially informed dispersal (Clobert *et al.* 2009), where individuals gather social information during their movements. Empirical evidence if this applies also during the transience phase is critically lacking. The ability of dispersers to assess the topology of the social landscape influences how and where they move, with possible consequences on the settlement process. The latter can in turn influence the dynamic and persistence of structured populations. While we are only scratching the surface in understanding the proximate mechanisms and demographic consequences of dispersal, our results are an important step towards a better understanding of socially informed dispersal in a spatially structured species.

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DATA ACCESSIBILITY STATEMENT Data supporting the results of this manuscript is archived in the public repository Dryad (Cozzi *et al.* 2018; <https://doi.org/10.5061/dryad.dk2pm>).

AUTHORS' CONTRIBUTION GC, AO and NM conceived the study, NM collected data on dispersing groups, TCB led the collection of data on resident groups, GC and LB performed analysis, GC wrote the first version of the manuscript, and all authors contributed substantially to revisions.

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Figure 1: Environmental and social geographic layers for the study area along the fossil Kuruman River, South Africa. The site was divided in five discrete habitat types subsequently merged into three: pans, transition and sandy matrix (A). For comparison, Google Earth images have been inserted (framed inserts). A digital elevation model for the study area ranging between 890 and 960 meters above sea level was derived from ASTER imageries (B). A kernel utilization density (KUD) map for each resident group was calculated reducing the reference smoothing parameter h_{ref} until the smallest possible contiguous territory (grey polygon) was obtained. Dots represent collected GPS locations (C1). KUD pixel values were then multiplied by group size. Light (yellow) colours represent frequently used regions (C2). KUD raster maps of each resident group were summed together to create a social landscape comprising all resident groups (C3). Example trajectories of a dispersing coalition that settled (dark dotted line) and a dispersing coalition that returned to the natal group (grey dashed line) are shown (C3). In all panels, the main river channel (solid blue line) are depicted for spatial reference.

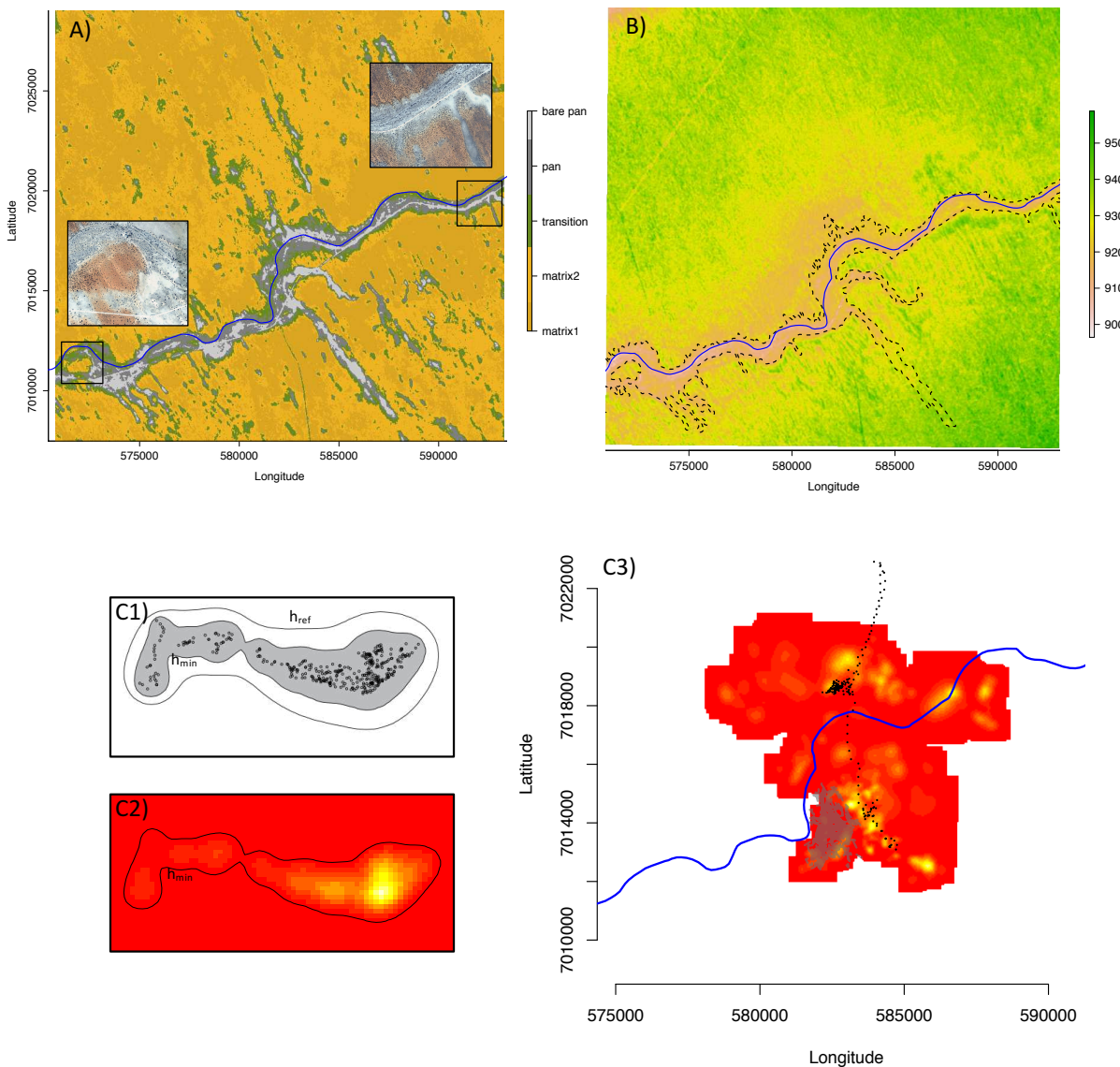


Figure 2: Dispersing coalitions step selection coefficients associated with the distribution of resident groups. The beta coefficients show the departure from baseline social landscape values that dispersing coalitions select when moving within the territory of the natal group. This means that when dispersing coalitions are outside the territory of the natal group they select for areas characterized by a lower likelihood of encountering resident conspecifics. Discrete social landscapes (real scenario) were calculated using GPS relocation data collected on resident groups during alternative time intervals (30, 45, ..., 240 days) prior to emigration by the dispersing coalition of interest. For comparison, a temporal mismatch between the social landscape and the dispersal events was created by shifting the time intervals used to create the social landscape back in time by 180 days (mismatch scenario).

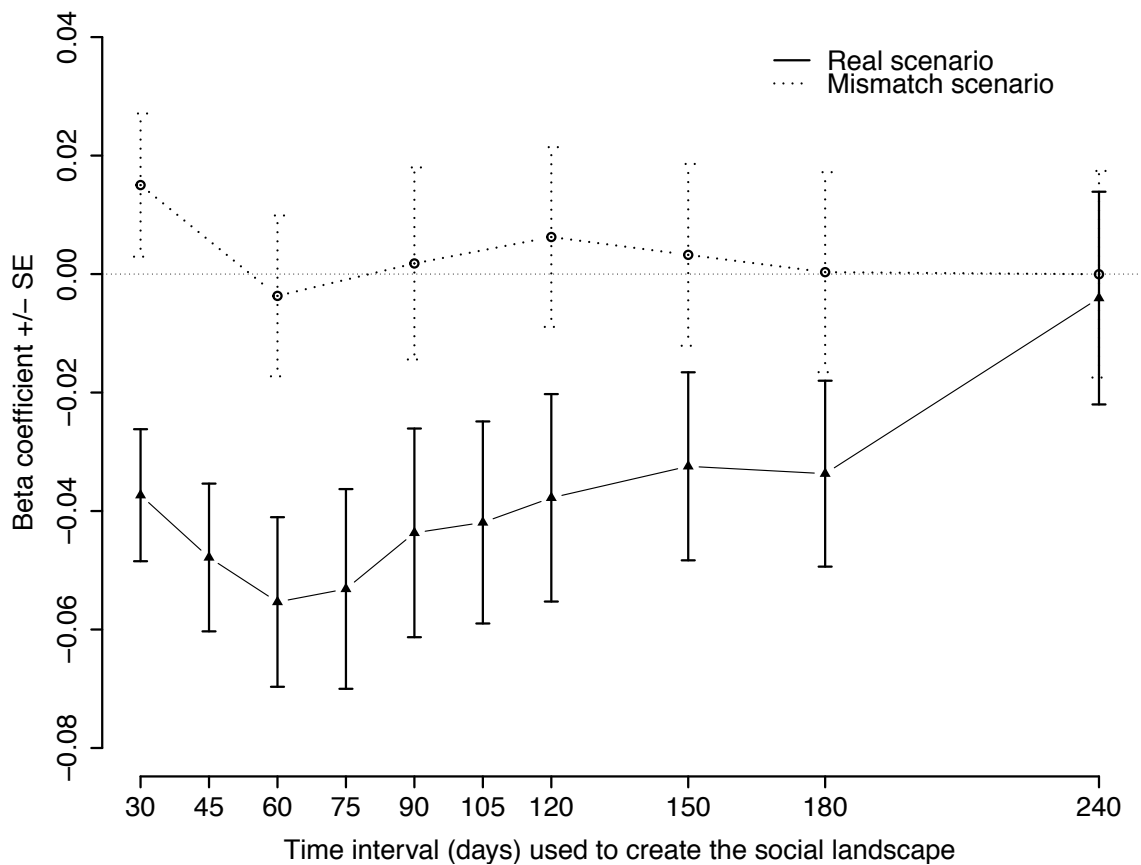


Figure 3: Coalition level relationship between selection for risky areas and dispersing coalition size. Beta coefficients were obtained from SSF fitted to single coalitions and based on data collected while the coalitions were within (A) and outside (B) the territory of the natal group. Removal of the apparent outlier in (B) did not change the results. Note that Y-axes in (A) and (B) are on different scales.

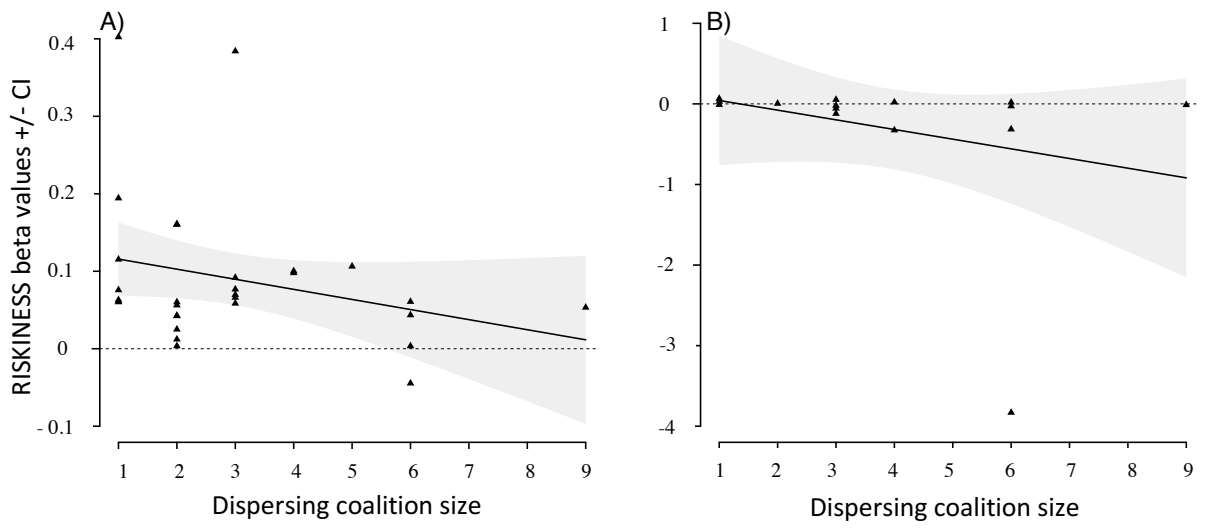


Table 1: Population-level selection coefficients. Coefficients and robust standard errors (S.E.) from a step selection function analysis where each observed location was conditioned on 10 alternative locations. Pans: factorial (whether a location is inside a pan or in the surroundings). Natal: factorial (whether a coalition is inside or outside the territory of the natal group). Riskiness: continuous values associated with the social landscape (range: $x \rightarrow 0$ to $\max(x)$). Elevation and distance to river are measured in meters. Coal size: Maximum coalition size. Removal of distance to river did not change the results.

	coefficient	S.E.	p-value
ELEVATION	-0.020	0.008	0.011
PANS (inside)	-0.259	0.131	0.049
DISTANCE TO RIVER	0.000	0.000	0.440
NATAL (outside)	-0.314	0.163	0.163
RISKINESS	0.059	0.009	< 0.01
NATAL(outside):RISKINESS	-0.055	0.014	< 0.01
COAL SIZE:RISKINESS	-0.005	0.003	0.090

Table 2: Individual-level selection coefficients. Coefficients and standard errors of movement steps during dispersal regressed against environmental and social variables. For a description of variables see Table 2.

Inside NATAL						Outside NATAL					
COALITION ID	COAL SIZE	ELEVATION	PANS (inside)	DISTANCE TO RIVER	RISKINESS	COALITION ID	COAL SIZE	ELEVATION	PANS (inside)	DISTANCE TO RIVER	RISKINESS
β						β					
1	1	-0.085	-16.354	-0.004	0.402	1	1	-0.093	-17.523	0.000	0.065
2	2	-0.037	-15.668	0.002	0.056	4	6	-0.025	0.390	0.000	-0.315
3	3	0.025	-16.995	-0.003	0.066	5	6	-0.003	-1.579	0.000	-0.027
4	6	-0.012	0.208	-0.004	0.043	8	2	-0.001	0.403	0.000	0.004
5	6	0.040	-1.583	0.000	0.003	31	1	-0.044	0.179	0.000	0.029
6	4	0.065	-0.112	0.000	0.100	10	6	-0.061	-0.163	0.000	0.019
7	3	0.103	-1.643	0.001	0.384	32	4	-0.013	-0.867	0.000	0.021
8	2	-0.027	0.925	0.000	0.042	33	3	-0.065	0.323	0.000	-0.021
9	1	0.005	-0.962	0.000	0.076	34	4	-0.018	0.093	0.000	-0.326
10	6	-0.010	-0.235	0.000	0.061	16	3	0.029	-0.092	0.000	0.053
11	2	-0.083	-17.720	0.000	0.160	35	6	-0.102	-0.354	0.000	-3.831
12	4	-0.028	-0.707	0.000	0.098	17	3	0.061	-1.244	0.001	-0.056
13	2	-0.086	-0.855	0.002	0.012	19	9	0.030	1.082	-0.001	-0.012
14	2	-0.005	-0.037	0.001	0.060	22	3	-0.166	-0.059	0.002	-0.121
15	5	-0.032	0.027	0.000	0.106	30	1	-0.068	-0.490	0.000	-0.010
16	3	-0.018	0.210	0.001	0.077						
17	3	-0.037	-0.407	-0.001	0.070						
18	3	-0.078	-0.120	0.000	0.092						
19	9	0.007	0.105	0.000	0.053						
20	2	-0.042	0.096	0.001	0.043						
21	1	-0.043	-1.288	0.002	0.194						
22	3	-0.037	-0.869	0.002	0.058						
23	2	-0.160	0.001	0.001	0.025						
24	2	0.139	-1.251	0.001	0.161						
25	1	0.044	0.268	-0.001	0.115						
26	1	-0.070	-0.283	0.000	0.060						
27	6	0.064	-0.326	0.000	-0.045						
28	2	-0.178	1.088	-0.001	0.004						
29	1	0.022	0.337	0.000	0.063						
30	1	0.012	0.101	-0.001	0.061						